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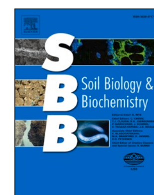

Eukaryotes in soil aggregates across conservation managements: Major roles of protists, fungi and taxa linkages in soil structuring and C stock

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ABSTRACT

The stabilization of soil organic carbon (SOC) promoted by conservation agriculture (CA) depends on soil aggregation. Aggregation protects SOC and creates heterogeneous microhabitats hosting diverse soil biota which in turn promote aggregation. A long-term experiment, studying the interaction of tillage with nitrogen (N) fertilization on a soybean-wheat rotation, was used to investigate eukaryotic community diversity, composition, and structure within small macroaggregates (sM) and occluded microaggregates (mM). Using high-throughput Illumina sequencing, we found (i) a different eukaryote diversity response to management intensification across soil aggregates and soil depths; (ii) a conserved core community composition of eukaryotes across CA treatments and aggregates at surface and subsurface layers; (iii) a different effect of tillage on eukaryotic community structure in sM and mM along the soil profile according to N availability; (iv) a positive association of protists, and fungi with the amount of sM and mM, and their SOC content; (v) a stronger complexity of within- and cross-domain networks (eukaryotes and eukaryotes-prokaryotes) in mM than in sM at surface layer. Overall, our findings demonstrate for the first time that protists together with fungi play major roles in soil structuring and C cycling, and that Cercozoa represent hubs in soil biota aggregate networks.

1. Introduction

Soil organic carbon (SOC) stability mainly depends on physical protection (Six et al., 2000; Six and Paustian, 2014), whereas molecular structure of plant residues and root exudates play a secondary role in SOC persistence (Schmidt et al., 2011; Lal et al., 2015). Organic carbon (C) is protected in soil aggregates by physically limiting the access of decomposers and enzymes and the diffusion of O_2.

According to the model of Tisdall and Oades (1982), primary particles (clay and silt particles, Ø <53 μm) are bound together by persistent bacterial, fungal, and plant debris into free microaggregates (Ø 53–250 μm). Free microaggregates are bound into macroaggregates (Ø > 250 μm) by transient agents (i.e., microbial and plant polysaccharides) that are rapidly decomposed by microorganisms, and by temporary agents (i.e., roots, fungal hyphae and glomalin) that persist in the medium term. Labile SOC is mainly located in macroaggregates, while free microaggregates contain a more recalcitrant SOC pool (Elliott, 1986; Jastrow and Miller, 1998).

Intensive agricultural practices, such as tillage and fertilization, shorten the life cycle of macroaggregates and diminish the formation rate of new microaggregates, worsening soil structure (Six et al., 2000). In no-tillage systems (NT), the slower turn-over of macroaggregates resulted in more sequestration of crop-derived C in microaggregates formed within macroaggregates (occluded microaggregates, mM; Ø 53–250 μm), and thus the amount of mM is crucial for the long-term C-sequestration in soils (Six et al., 2000; Denef et al., 2007; Sheehy et al., 2015). In this context, the application of conservation agriculture (CA) practices (i.e., minimum tillage/NT, crop rotation and mulching) may allow the establishment of microhabitats with variable nutrient availabilities for a diverse soil biota, acting as efficient binding agent (Kong et al., 2011; Gupta and Germida, 2015; Totsche et al., 2018; Piazza et al., 2019). Moreover, CA practices may also produce yields equivalent to or even greater than conventional systems (Rusinamhodzi et al., 2011; Aune, 2012; Pittelkow et al., 2015; Himmelstein et al., 2016).

In boreal climates, long-term NT and minimum tillage (MT) have been shown to increase the amount of macroaggregates and mM as well as their SOC content in the shallow layer (surface soil within horizon A)

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in comparison to conventional tillage (CT) (Franzluebbers and Arshad, 1997; Sheehy et al., 2015). This was demonstrated under various soil textures and was more evident in clay, clay-loam and silt-loam soils. Similarly, in humid tropical climates and sandy loam soils, long-term application of MT significantly increased SOC content in large soil aggregates, whereas the reverse was reported under CT (Onweremadu et al., 2007). Accordingly, Denef et al. (2007) highlighted a promotion of mM fraction and mM-associated C stocks in NT and MT compared with CT under similar climate. Moreover, nitrogen (N) fertilization was reported to increase SOC in macroaggregates and free microaggregates by decreasing the activity of cellulolytic fungi and bacteria (Ghosh et al., 2019; Duan et al., 2021). Recently, in a cold and humid Mediterranean area and in a silt-loam soil, high N fertilization rates in combination with MT not only increased mM, but also promoted a shift to low level, but more efficient C-cycling microbial enzyme activities, which were correlated to a greater accumulation of SOC (Piazza et al., 2020). Overall, in four regions across Europe the intensification of agriculture was reported to consistently reduce soil biota diversity in bulk soil, making soil food webs less diverse and composed of smaller bodied organisms (Tsiafouli et al., 2015).

Although the role of bacterial and fungal communities (including arbuscular mycorrhizal fungi, AMF) in soil aggregation and SOC stabilization is widely recognized to be fundamental (Six et al., 2004; Lehmann et al., 2017; Bach et al., 2018), the diversity and potential role of other soil biota have received less attention. Soil biota diversity has indeed proved to be the major driver of C sequestration and nutrient cycling in bulk soil (De Vries et al., 2013; Wagg et al., 2019; Delgado-Baquerizo et al., 2020). Many studies demonstrated that earthworms and bacterivore nematodes are directly involved in the formation of macroaggregates by incorporating fresh organic matter inside mM and thus promoting SOC accumulation (Six et al., 2004; Pulleman et al., 2005; Bossuyt et al., 2006; Fonte et al., 2007; Zhang et al., 2013; Delgado-Baquerizo et al., 2020). Moreover, an indirect effect on SOC accumulation by earthworms and bactrivorous nematodes was also reported and explained by the shift of soil microbial diversity through taxa regulating nutrient flow (Delgado-Baquerizo et al., 2020). Thus, in this study, we investigated the diversity and related roles of the eukaryotic component of soil biota within small macroaggregates (sM) and mM across CA managements. Moreover, since a more connected soil biota network takes up more C (Morriën et al., 2017), the study was extended to elucidate how eukaryotes are connected among each other, and to the prokaryotic community. In this context, long-term CA field experiments in the Mediterranean area, such as the one used in this study, provide a great opportunity for improving the understanding of soil eukaryotic diversity and functionality in soil aggregates and C stocks.

The following hypotheses were tested: (1) conservation tillage and N fertilization shift soil eukaryote community diversity, composition and structure, in soil aggregates along the soil profile; (2) soil aggregates differentially shape the diversity, composition and structure of soil eukaryotes; (3) some eukaryotic taxa are predictors for soil structuring and C stocks; (4) eukaryotes form structured assemblages and distinctive networks in soil aggregates (within-domain networks); (5) the traits of the eukaryotes-prokaryotes networks vary across aggregates (cross-domain networks), and some network traits can predict soil structuring and C stocks.

2. Materials and methods

2.1. Field experiment

A long-term CA field experiment on a bread wheat (*Triticum aestivum* L.) - soybean (*Glycine* max L. Merr.) rotation was set up in 1993 at the Centro Interdipartimentale di Ricerche Agro-Ambientali Enrico Avanzi (Pisa, Italy; 43°40′ latitude N; 10° 19’ longitude E; 1 m above sea level) in an alluvial silt loam soil (131, 613 and 256 g kg^{-1} of sand, silt and clay, respectively). The experiment was conducted comparing two tillage intensities and two N fertilization levels. The tillage intensities were: conservation tillage (minimum tillage, MT: disk harrowing at 15-cm depth) and conventional tillage (CT: mouldboard ploughing at 25-cm depth, disking and harrowing at 15-cm depth). The N fertilization levels applied only to bread wheat were: 0 and 200 kg N ha^{-1} (N0 and N200, respectively). The soil is classified as Typic Xerofluvent by USDA system (Soil Survey Staff, 1975) and as Fluvisol by FAO (IUSS working group WRB, 2006). Climate of the site is cold, humid Mediterranean (Csa), according to the Köppen-Geiger climate classification (Kottek et al., 2006). The experiment was arranged following a split-plot design, with tillage as main-plot factor and N fertilization as subplot factor and three replicate plots (dimension: 11.5 × 14.5 m). The N fertilizer treatment was applied as urea and the rate was split into three applications, before sowing (60 kg N ha^{-1}), at the first detectable node (70 kg N ha^{-1}), and 15 days after this stage (70 kg N ha^{-1}). Under CT, almost 100% of the residues were incorporated in the 0–25 cm soil layer, whereas under MT approximately 50% of the crop residues were incorporated at 0–15 cm depth. Crops were managed applying pre-emergence herbicide for weed control and no disease or insect treatments (Piazza et al., 2020).

2.2. Soil sampling and analysis of soil physical and chemical parameters

Soil sampling was carried out in Spring 2016) before soybean sowing. In each replicate plot, a homogenized sample was obtained by mixing four soil cores collected at two soil depths (surface layer: 0–15 cm; sub-surface layer: 15–30 cm). A total of twenty-four soil samples were collected (12 at the surface layer; 12 at the subsurface layer). Once in the laboratory, each sample was air-dried, gently broken apart and then passed through an 8-mm sieve. The isolation of small macroaggregates (sM; 250–2000 μm) was done from 80 g of the sieved soil samples by the wet sieving method (Six et al., 1999). Occluded microaggregates (mM; 53–250 μm) were isolated from an additional isolation of sM (i.e., starting from 80 g of the sieved soil samples) and utilising a device designed and built by Piazza et al. (2020). Once collected, the fractions were freeze-dried (FreeZone 2.5 Labconco, Kansas City, MO, USA) for 48–72 h for dry weight determination and chemical and molecular analyses. Both aggregate fractions of all samples were then analysed for SOC by CHN combustion method (LECO, Italy) and SOC content was calculated and expressed in Mg ha^{-1} (Bremner and Mulvaney, 1982; Piazza et al., 2020).

2.3. Molecular analyses

DNA was extracted from 0.25 g of sM (n = 24) and mM samples (n = 24) using the DNeasy PowerSoil Kit (QIAGEN, Venlo, Netherlands). The DNA extracts were then quantified by a spectrophotometer (NanoDrop Technology, Wilmington, DE) and stored at −20 °C. PCRs were generated from 10 ng μL^{-1} genomic DNA in volumes of 25 μL with 0.125 U μL^{-1} of GoTaq® Hot Start Polymerase (Promega Corporation, WA, USA), 0.5 μM of each primer, 0.2 mM of each dNTP, 1 mM of $MgCl_2$ and 1x reaction buffer, using the PTC-200 96-well Peltier Thermal Cycler (MJ Research, MA, USA). The primers were TAReuk454FWD1-ill (5′-TCGTCGGCAGCGTCAGATGTGTATAAGAGACAGANNNHNNNWNNNH CCAGCASCYGCGGTAATTCC-3′) and TAReukREV3-ill (5′- GTCTCGTG GGCTCGGAGATGTGTATAAGAGACAGTACTTTCGTTCTTGATYRA-3′) (modified from Stoeck et al., 2010). The primer pair has attached Illumina sequencing tags, and for the forward primer a 13 bp random sequence was included in order to improve cluster definition on the MiSeq slide. Primers target the hypervariable region V4 of the small subunit ribosomal RNA (SSU rRNA 18S) gene fragment. The thermal cycler was programmed as follows: 95 °C for 2 min, 35 cycles at 94 °C for 30 s, 50 °C for 45 s, 72 °C for 1 min and 30 s and a final extension step at 72 °C for 10 min. PCR products were examined by electrophoresis through a 1% agarose gel in 0.5 × TBE buffer, then purified with magnetic beads (Agencourt® AMPure® XP, Beckman Coulter, USA) and freshly prepared 80% ethanol, and quantified by fluorimetry with the

use of Quant-iT™ dsDNA HS (High-Sensitivity) Assay Kit (Invitrogen by Thermo Fisher Scientific, CA, USA), following the instructions of the manufacturer. Cleaned and quantified amplicons of each library were adjusted in an equimolar ratio (10 ng μL^{-1}) for the required Illumina P5 and P7 sequences addition along with index sequences in a new PCR step. Indexing was performed using primers from the Nextera® Index kit (sets A and D; Illumina Inc., CA, USA) and the resulting metabarcoding libraries were sequenced on an Illumina MiSeq sequencer (2 * 300 bp paired-end reads) at the Genomics and Bioinformatics Laboratory (Technology Facility, Department of Biology, University of York, UK). Details are given in the Supplementary Methods 1.

2.4. Bioinformatic analyses

Raw data generated from the Illumina MiSeq sequencing run were processed and analysed following the pipelines of QIIME 2 (2018.4) and USEARCH (v10.0.240) (Edgar, 2010; Caporaso et al., 2012). Forward and reverse paired-end sequences were assembled independently for each sample using -fastq_mergepairs USEARCH command. Primer sequences were then trimmed off by employing cutadapt plugin (2018.4) with default settings. To avoid potential errors in sequencing data, quality of sequence reads was checked by -fastq_eestats2 USEARCH command, using the expected number of errors in a read as a measure of quality for filtering (Edgar and Flyvbjerg, 2015). Reads were then trimmed at the length where the “drop-off” point for the maximum expected error value occurred (250 bp). Quality filtered reads were de-replicated by -fastx_uniques USEARCH command, then Operational Taxonomic Units (OTUs) were generated using USEARCH by clustering sequence reads at the 97% similarity threshold. During the process, chimeric sequences and singletons were removed from the dataset. For the curation, the sequences were aligned using ClustalW and then Neighbor Joining (NJ) phylogenetic tree was built in MEGA7 (Kumar et al., 2016) (https://www.megasoftware.net). The most abundant sequence of the eukaryotic OTU in each cluster was selected, and used as representative sequence for that OTU after branch collapsing. For the curation, the sequences were aligned using ClustalW and then Neighbor Joining (NJ) phylogenetic tree was built in MEGA7 (Kumar et al., 2016) (https://www.megasoftware.net). The most abundant sequence of the eukaryotic OTU in each cluster was selected, and used as representative sequence for that OTU after branch collapsing. The OTUs were phylogenetically assigned using the 18S SSU SILVA database (version 132, release date December 13, 2017) (Quast et al., 2012; Yilmaz et al., 2013) by clustering sequence reads at the 97% similarity threshold. After curation, the representative sequences were re-aligned using ClustalW and the phylogenetic tree was built in MEGA7 using the Neighbor Joining (NJ) analysis with 1000 bootstrap replicates and the Kimura 2-parameter model (uniform rates).

The suitability of the eukaryotic community sampling was verified by rarefaction curves plotting the number of eukaryotic classes/phyla *versus* the number of sequence reads, while accumulation curves were calculated plotting the number of classes/phyla *versus* the number of soil samples using the package Vegan in R (Oksanen et al., 2013). Since there was a high variability in the number of reads per sample, sequencing depth per sample was standardized to the median number of reads across the samples in each data matrix using the same package in R (standardized datasets). All representative sequences were deposited in the NCBI GenBank database (SUB5948379 submission: MN178662-MN178794 accession numbers).

Prokaryotic data were obtained from the same soil matrices and depths (Piazza, 2019) and were based on the V4 region of the 16S rRNA gene sequenced using a MiSeq Illumina approach (SUB5941754 submission: MN171543-MN172157 accession numbers).

2.5. Statistical analyses

To test hypothesis 1 - Conservation tillage and N fertilization shift soil eukaryote community diversity, composition and structure in soil aggregates along soil profile - analysis of variance (ANOVA), Venn diagrams and permutational analysis of variance (PERMANOVA) were applied. Concerning diversity, richness, Shannon index (H') and Simpson index ($\lambda = 1 - \lambda'$) were calculated at class level and analysed by two-way ANOVA, according to the experimental design. These analyses were done in Vegan package in R and plotted by ggplot2 (Wickham and Chang, 2008). Data were ln–transformed when needed to fulfil the assumptions of ANOVA. Post-hoc Tukey-B significant difference test was used for comparison among treatments. Concerning composition, Venn diagrams were drawn to visualize the OTUs unique to the treatments as well as the shared ones. The standardized datasets were used to generate the Venn diagrams by the online tool InteractiVenn (http://www.interactivenn.net; Heberle et al., 2015). Concerning structure, the relative abundances of eukaryotes were calculated at class level and the permutational analysis of variance (PERMANOVA) (Anderson and ter Braak, 2003) and the analysis of homogeneity of multivariate dispersion (PERMDISP) were used to test the effect of treatments (Clarke and Gorley, 2006; Torgerson, 1958). Response data were square-root transformed to down-weight the dominant taxa and the Bray-Curtis index of dissimilarity was used to measure ecological distance. When PERMANOVA indicated a significant effect, the principal coordinate analysis (PCO) was carried out (Anderson et al., 2008) to visualize the most relevant patterns in the response data. In each PCO biplot, only the taxa with a strong correlation ($r = 0.50$–0.80) with the ordination scores on each PCO axis were displayed. *P*-values were calculated using the Monte-Carlo test and residuals were permuted according to the experimental model (Oksanen et al., 2013). Multivariate analyses were performed using PRIMER 6 and PERMANOVA + software (Clarke and Gorley, 2006; Anderson et al., 2008).

To test hypothesis 2 - Soil aggregates differentially shape the diversity, composition and structure of soil eukaryotes - eukaryotic richness, H' and λ at phylum level and at both soil depths were analysed by one-way ANOVA, using soil matrix (sM *vs* mM) as fixed factor, and tillage and N fertilization as covariates. Analyses were performed in Vegan package in R and data were plotted by ggplot2. Data were ln-transformed when needed to fulfil the assumptions of ANOVA, and the post-hoc Tukey-B significant difference test was used for comparison among treatments. Moreover, the effect of matrix on composition and structure were analysed at phylum level using the Venn diagrams and PERMANOVAs, as described above.

To test hypothesis 3 **-** Some eukaryotic taxa are potential predictors for soil structuring and C stocks - multiple regression analysis was applied using as independent variables the standardized relative abundances (calculated as described above) of eukaryotic taxa at class level. To account for the compositional nature of the data, an additive log-ratio transformation was applied (Gloor et al., 2017). The dependent variables were sM and mM weights, and SOC content per unit of surface in sM and mM. The assumptions of the linear regression model were verified (Supplementary Method 2) and the multiple linear regression analysis was applied using a stepwise method with the following probability criteria: $P < 0.05$ to accept and of $P > 0.05$ to remove a phylum or a within/cross-domain network traits. Multiple regressions were performed using the SPSS software package version 25.0 (SPSS Inc., Chicago, IL, United States of America)**.** Details about regression analysis are reported in Supplementary Methods 2.

To test hypothesis 4 - Eukaryotes form structured assemblages and distinctive networks in soil aggregates - we built networks using the SParse InversE Covariance estimation for Ecological ASsociation Inference (SPIEC-EASI) package version 0.1 in R (https://github.com/zdk123/SpiecEasi/). SPIEC-EASI is a pipeline for inferring sparse inverse covariance matrix within and between multiple compositional datasets, under joint sparsity penalty (Kurtz et al., 2015). The within-domain analyses were performed on the standardized eukaryotic dataset at class level for each soil matrix and depth. The neighborhood selection (MB method) was applied as graphical inference model

(Meinshausen and Bühlmann, 2010), since it has been shown to better perform than other available methods (e.g., CCREPE, SPARCC, SPIEC-EASI glasso) (Kurtz et al., 2015). The Stability Approach to Regularization Selection (StARS) was applied to select the optimal sparsity parameter (Liu et al., 2010), and the StARS variability threshold was set to 0.05 and nλ to 100 for all networks. We evaluated the weights of the edges in the networks using SPIEC-EASI (frequency *versus* edge weights = modularity), and we plotted the degree distributions of frequencies of the edges using adj2igraph (Kurtz et al., 2015). In the networks a node represents a connected taxon, an edge the connection between taxa, a singleton an unconnected taxon and a dyad two connected taxa. For the eukaryotic networks (within-domain network) we calculated the following parameters: number of nodes excluding singletons, number of edges, number of singletons and dyads, number of subnetworks (a subnetwork is a network composed by at least three nodes), mean nodes per subnetwork, linkage density (complexity: the average number of edges per node), percentage of positive interactions and modularity. Moreover, in each network, we calculated the frequency of the phyla within the subnetworks, the mean of edges and nodes, and the percentage of positive edges for each phylum. Details about trait calculations are given in Supplementary Methods 3.

To test hypothesis 5 - The traits of the eukaryotes-prokaryotes networks vary across aggregates we inferred the associations between eukaryotes and prokaryotes domains by the cross-domain extension of SPIEC-EASI (Kurtz et al., 2015; Tipton et al., 2018). The same traits calculated for within-domain networks, except for the positive edges, were assessed (Supplementary Methods 3). In addition, the percentage of eukaryotes/prokaryotes per subnetwork was calculated and the number of subnetworks with only eukaryotes, only prokaryotes and with both domains were counted. The significance of the cross-domain relationships was tested by the Mantel test (Mantel and Valand, 1970) on the standardized read data that were centered and normalized and using the function Jaccard in PC-ORD 5 to build the resemblance matrix (Grandin, 2006) and also by the co-Correspondence Analysis (CoCA) (ter Braak and Schaffers, 2004) in CANOCO 5 (ter Braak and Smilauer, 2012). To test if some network traits can predict soil structuring and C stocks - a multiple linear regression was performed after verification of the assumptions (Supplementary Methods 3). The analysis was performed using as independent variables the log(1+x)-transformed and normalized within- and cross-domain network traits in sM and mM (i.e., traits of eukaryotic networks and of eukaryotic-prokaryotic networks), and using as dependent variables sM and mM weights, and SOC content per unit of surface in sM and mM. Scripts for within and cross-domain network construction and analysis are available in Supplementary Methods 4.

3. Results and discussion

3.1. Illumina sequencing information

MiSeq sequencing yielded a total number of 2 940 322 reads from the 48 soil samples, and following quality-filtering a total of 2 884 052 sequence reads having a length of 392 bp were obtained. After BLAST against the 18S SSU SILVA database (Quast et al., 2012; Yilmaz et al., 2013), we found 2 036 277 reads, ranging from 3 to 106 540 reads per sample that were assigned to a total of 4211 OTUs. After sequence curation and the removal of Plantae sequences, 863 809 reads, ranging from 6421 and 46 429 reads per sample, were retrieved and assigned to 133 OTUs, 56 classes and 27 phyla (Fig. S1, Fig. S2). The rarefaction and accumulation curves demonstrated that sampling effort was sufficient as the curves reached the asymptote (Fig. S3, Fig. S4).

3.2. Effect of conservation management on eukaryotic diversity in soil aggregates

To test if conservation tillage and N fertilization shift the diversity of soil eukaryotes in soil aggregates (hypothesis 1), richness and diversity indices were determined along the soil profile in small macroaggregates (sM) and occluded microaggregates (mM). A greater eukaryotic diversity in sM was consistently found at both soil layers under CT compared to MT, as shown by the significant increase of richness (+26%) and *H'* and λ (+9%) (Fig. S5). This higher eukaryotic diversity might be due to larger root development and higher availability of root exudates, organic matter (e.g., nutrients, organic acids), water and oxygen, reported under deep ploughing systems, and which have been shown to promote microbial growth and soil biota diversity/functionality (Guan et al., 2014; Edwards et al., 2015; Ercoli et al., 2017; Piazza et al., 2020), according to the response of individual taxonomic units to habitat and trophic conditions (van Capelle et al., 2012). At surface layer, N fertilization significantly increased λ under MT, suggesting an increase in number of relative abundances of taxa regulated by N availability. Conversely, under CT, λ was high and not modified by N fertilization, suggesting non-limiting N availability due to improved plant growth and higher mineralization rate of residues. A low eukaryotic diversity was previously found in bulk soil under N fertilization (Lentendu et al., 2014), and within soil aggregates a higher microbial diversity was found at low nutrient availability and NT compared with high nutrient availability and ploughing (Lagomarsino et al., 2012; Zhang et al., 2013; Bach et al., 2018).

In mM an opposite pattern was found at surface layer, and no effect of N fertilization alone or in interaction with tillage was observed at both soil depths (Fig. S5). At surface layer, the eukaryotic diversity indices increased by 6% under MT compared to CT. Under tillage intensification, macroaggregates are indeed disrupted and occluded microaggregates became free in the soil (Six et al., 2000), potentially reducing diversity and C sequestration. This is supported by the highest SOC accumulation observed in mM under MT (Piazza et al., 2020), and by the high soil biota diversity found in the present study within mM.

To test if soil aggregates differentially shape the diversity of soil eukaryotes (hypothesis 2), richness, *H'* and λ were determined in sM and mM. Overall eukaryotic diversity was significantly higher in mM than in sM (Fig. S5) (i.e., at surface layer: richness +16%, *H'* +6%, and λ +2%; at subsurface layer: richness +20 and *H'* +5%). These results are in accordance with the higher richness and *H'* of bacteria and fungi found in free microaggregates compared to large macroaggregates (Bach et al., 2018). Our findings support the fact that soil aggregates are distinct habitat spaces with eukaryotes adapted to SOM resources, pore-space network, and water and oxygen availability characteristic of sM and mM.

3.3. Effect of conservation management on eukaryotic composition in soil aggregates

To test if conservation tillage and N fertilization shift the composition of soil eukaryotes in soil aggregates (hypothesis 1) and how soil aggregates shape their composition (hypothesis 2), this parameter was evaluated along the soil profile in sM and mM. Across management practices and soil depths, the eukaryotic phyla Cercozoa (21%), Ciliophora (13%), Chlorophyta (11%), Nematoda (11%) and Glomeromycota (9%) were predominant in sM, whereas in mM the predominant phyla were Ciliophora (19%), Cercozoa (18%) Chlorophyta (15%), and Ascomycota (11%) (Fig. 1a). The other phyla showed an abundance ≤8%. Sun et al. (2021) found that protists were the most dominant eukaryote (33.9% of the total eukaryotic sequences) in bulk soil. By contrast, Treonis et al. (2018) analysing the whole eukaryotic structure in bulk soil found a high abundance of fungi, Arthropoda, Nematoda and Anellida (40, 20, 20 and 11%, respectively), and a low abundance of protists (0.63%). Among protists, Rhizaria was the group with the highest relative abundance in arable soil, comprising as dominant taxa Cercozoa and Amoebozoa (Bates et al., 2013; Degrune et al., 2019a; Santos et al., 2020). Similarly, in another study, fungi were reported to be the most abundant (i.e., Ascomycota, Basidiomycota,

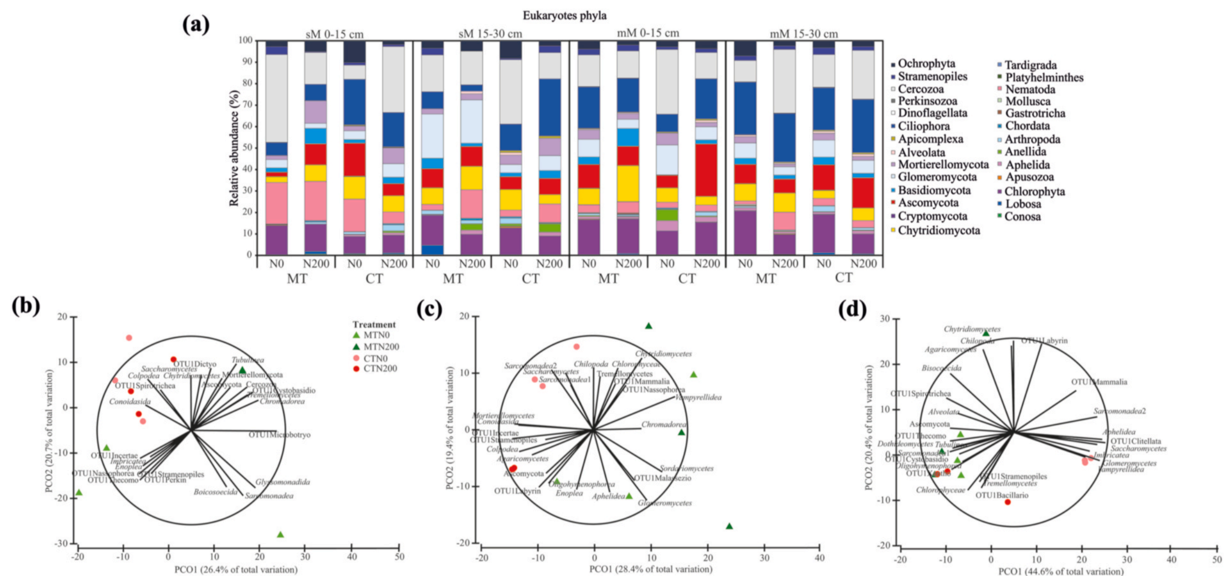

Fig. 1. Long-term effect of conservation management on community composition of eukaryotes in two soil matrices: small macroaggregates (sM) and occluded microaggregates (mM). Managements were: MTN0 (minimum tillage and 0 kg N ha^{-1}), MTN200 (MT and 200 kg N ha^{-1}), CTN0 (conventional tillage and 0 kg N ha^{-1}) and CTN200 (CT and 200 kg N ha^{-1}). Neighbor-joining (NJ) tree of 56 eukaryotic taxon representative sequences (classes) found in (a) sM and (d) mM. NJ trees are based on the sequences obtained from the amplification of the V4 region (18 SSU rRNA gene). The eukaryotic taxa were assigned to Operational Taxonomic Unit (OTU) (at class phylogenetic resolution) by BLAST against the 18S SSU SILVA database by clustering sequence reads at the 97% similarity threshold. For each OTU, the proportion of sequences retrieved in each management (MTN0, light green; MTN200, dark green; CTN0, light red; CTN200, dark red) and soil depth (0–15 cm: light grey; 15–30 cm: dark grey) are shown in the pie charts. Venn diagrams of eukaryotic classes uniquely retrieved and shared across managements in sM at (b) 0–15 cm and (c) at 15–30 cm, and in mM at (e) 0–15 cm and (f) at 15–30 cm.

fungi Incertae sedis and Glomeromycota), followed by Alveolata, Metazoa, Rhizaria, Stramenopiles, and Viridiplantae (Chen et al., 2012). Therefore, we can assume that the differences between bulk soil and soil aggregates depend on the variability of pH, moisture and organic nutrient availability that shift soil biota at multiple trophic levels.

In both soil aggregates, the majority of taxa were common to all managements across soil depths, whereas some were unique to certain managements, as shown by the Venn diagrams at class resolution (Fig. 2b–c,e-f). This is also shown by the pie charts representing the proportion of the 56 classes retrieved in each management (Fig. 2a,d). Moreover, focusing on the shared taxa between sM and mM, ca. 80% of eukaryotes were common to both soil aggregates, averaging soil depths (Fig. 3a and b). Accordingly, a large conserved core community of soil prokaryotes and fungi was found across managements at the same site in bulk soil (Piazza et al., 2019). This is also consistent with the findings obtained in other studies in different soil types and managements in bulk soil as well as in specific rhizo-compartments (Lentendu et al., 2014; Edwards et al., 2015; Pershina et al., 2018).

However, the exclusive presence of some eukaryotic taxa in the different systems and soil aggregates (Fig. 2) suggests that long-term tillage and N fertilization may drive the development of communities of specialized taxa putatively having specific functions (e.g., soil aggregate and/or SOC accumulation and nutrient cycling). As example, at the surface layer, in sM the classes *Nassophorea* and *Perkinsea* were exclusively found in MTN0, whereas *Pezizomycetes* and *Gastrotricha* in CTN0 (Fig. 2a and b), while many taxa were exclusively found in mM, such as Chlorophyta, *Eutardigrada*, *Eurotiomycetes*, *Nassophorea* and *Thecofilosea* in MTN0; Alveolata, *Chilopoda* and *Rhabditophora* in MTN200 and *Dictyostelia* and *Pezizomycetes* in CTN200 (Fig. 2d and e). A more in-depth description of the eukaryotic composition and exclusiveness across managements and aggregates is reported in the Supplementary Results and Discussion 1.

3.4. Effect of conservation management on eukaryotic community structures in soil aggregates

To test if conservation tillage and N fertilization shift the structure of soil eukaryotes in soil aggregates (hypothesis 1), the relative abundance pattern of taxa was determined in soil aggregates. Despite the high degree of similarity among treatments in term of community composition, we highlighted a strong effect of the interaction between tillage and N fertilization on the eukaryotic community structures in sM at both soil layers, and in mM only at the surface layer (Table 1, Fig. 1). Similarly, soil fungal community structure in bulk soil was strongly shaped by the interaction between tillage and N fertilization at surface and subsurface layers (Piazza et al., 2019). However, to our knowledge no studies have focused on the effect of the interaction of these practices on the eukaryotic communities in soil aggregates, whereas a huge number of studies was performed to assess the effect of tillage or N fertilization on the diversity/abundance and functionality of single eukaryotic group in bulk soil (e.g., fungi: Jansa et al., 2003; Wang et al., 2019; Zhao et al., 2019; micro-arthropods, nematodes and protozoa: Adl et al., 2006; Zhang et al., 2012; Briones and Schmidt, 2017; Cai et al., 2020; protists: Zhao et al., 2019; Sun et al., 2021).

In the PCO plots, CTN0 and CTN200 showed similar community structures within sM at surface layer (Fig. 1b) that were characterized by *Colpodea* and OTU1*Spirotrichea* (Ciliophora, Alveolata) and *Conoidasida* (Apicomplexa, Alveolata). This supports that under ploughing N availability is not limiting and the community structures are not affected by N fertilization. The class *Colpodea* is a well-known dominant clade of mainly bacterivorous protists (Foissner, 1998) and, according to our results, they were shown highly abundant in disturbed soils. By contrast, *Spirotrichea* were found in more stable environments (Lüftenegger et al., 1985), not supporting the large abundance found within sM under ploughed soils. This result not supported by literature may suggest a high resistance of *Spirotrichea* to natural and anthropogenic stresses. Moreover, Apicomplexa that are the third most abundant protistan group in soil, after Cercozoa and Ciliophora (Fierer, 2017) are described as putative parasites of invertebrates (Del Campo et al., 2019). This is in

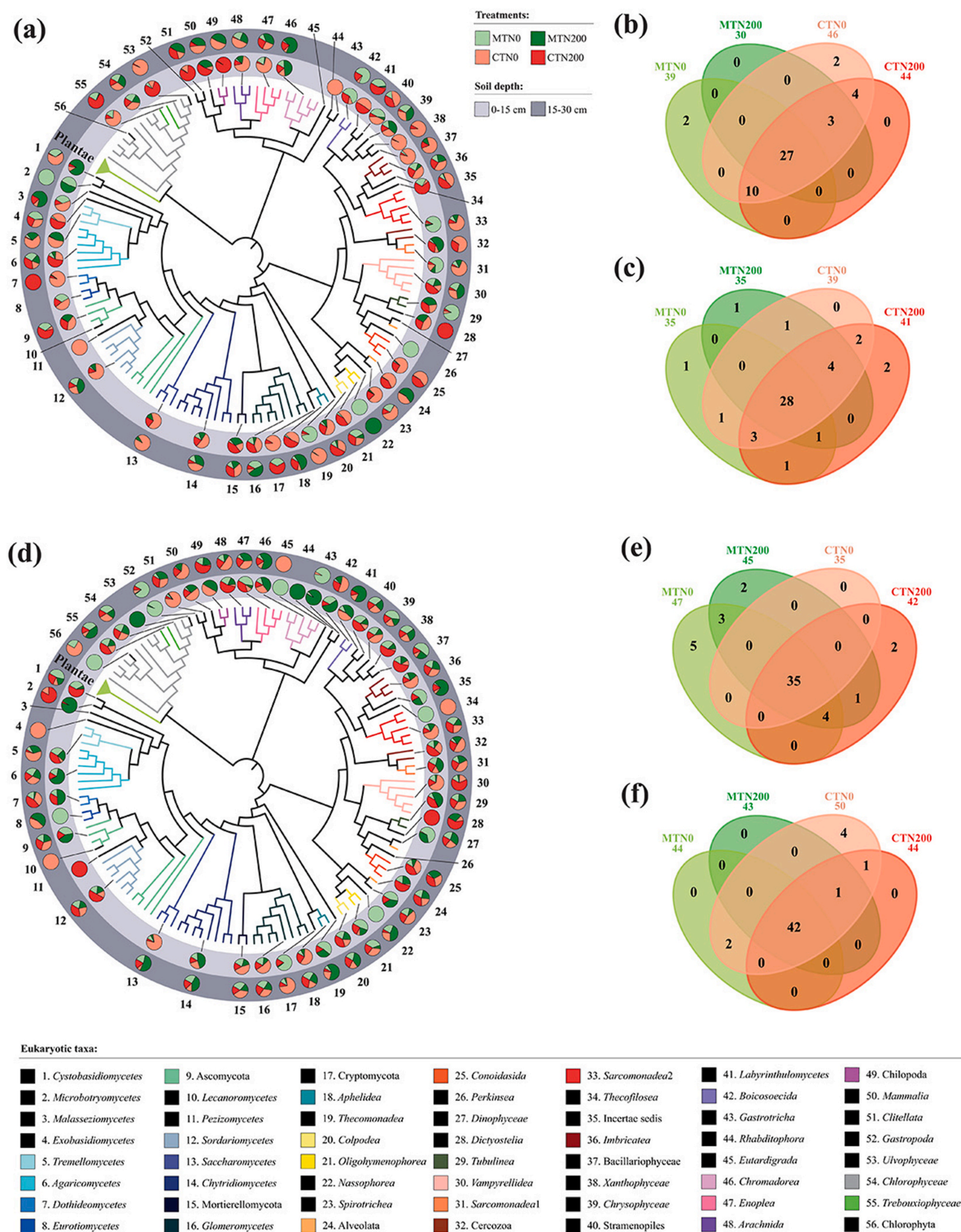

Fig. 2. Long-term effect of conservation management on relative abundances and community structures of eukaryotes in two soil matrices: small macroaggregates (sM) and occluded microaggregates (mM). Managements were: MTN0 (minimum tillage and 0 kg N ha^{-1}), MTN200 (MT and 200 kg N ha^{-1}), CTN0 (conventional tillage and 0 kg N ha^{-1}) and CTN200 (CT and 200 kg N ha^{-1}). (a) Relative abundances of eukaryotes at phylum level across treatments, matrices and soil depths. (b) Principal Coordinates Analysis (PCO) biplots on the interaction of tillage and N fertilization on the eukaryotic community structure at class level in sM at 0–15 cm and (c) at 15–30 cm, and (d) in mM at 0–15 cm. The output of the PCO biplots is based on the significant effect of treatments following the permutational analysis of variance (PERMANOVA). We displayed only the taxa with a strong correlation (r = 0.50–0.70) with the ordination scores on each PCO axis.

agreement with the lower occurrence of the invertebrates *Chromadorea* and *Enoplea* found in our study under CT compared with MT. In MT, N fertilization determined a strong shift at the surface layer (Fig. 1b), with community structure within sM under MTN200 characterized by a large abundance of taxa belonging to Cercozoa, *Tubulinea* (Lobosa, Amoebozoa) and *Chromadorea* (Nematoda) together with fungi, (i.e., OTU1Microbotryo: *Microbotryomicetes*, Ascomycota and Mortierellomycota), and under MTN0 characterized by a large abundance of taxa belonging to *Imbricatea* (Cercozoa), *Enoplea* (Nematoda) and OTU1*Nassophorea* (Ciliophora, Alveolata). The dominance of Ascomycota in sM under MTN200 is in accordance with their higher abundance in macroaggregates under mineral fertilization compared with no fertilization (Liao et al., 2018; Wang et al., 2021). This confirms the importance of fungi as binding agents in soil aggregates (Six et al., 2000). Cercozoa were reported to be affected by several abiotic factors, as soil moisture, clay content and N availability (Lentendu et al., 2014; Fiore-Donno et al., 2019). However, although our analyses did not allow discrimination of which cercozoan classes were favoured under MTN200 (Fig. 1b), the

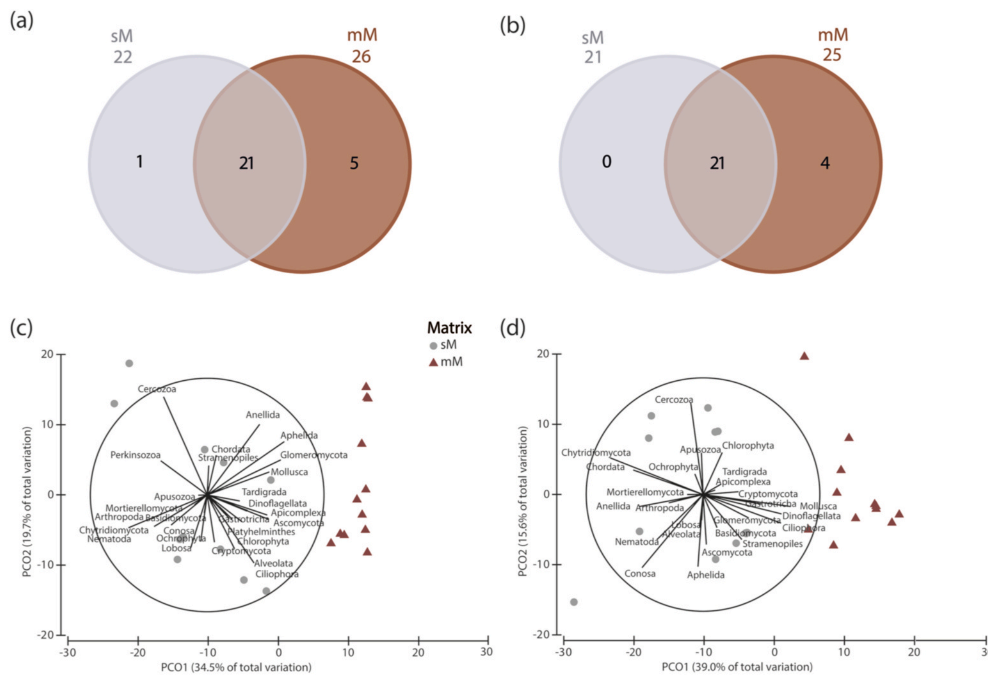

Fig. 3. Venn diagrams of eukaryotic phyla uniquely retrieved and shared in sM and mM at 0–15 cm (a) and at 15–30 cm (b). Principal Coordinates Analysis (PCO) biplots on effect of soil matrix (small macroaggregates sM *vs* occluded microaggregates mM) on eukaryotic community structure at phylum resolution at (a) 0–15 cm and at (b) 15–30 cm soil depth. The PCO biplots are based on the significant effect of soil matrix according to the permutational analysis of variance (PERMANOVA). In the biplots, only the taxa with a strong correlation (r = 0.50–0.70) with the ordination scores on each PCO axis were displayed.

Table 1
Results of PERMANOVA and variation partitioning of the long-term effect of conservation management (tillage and N fertilization) within small macroaggregates (sM) and occluded microaggregates (mM) and of the effect of matrix (sM *vs* mM) on eukaryotic community structure at 0–15 and 15–30 cm soil depths.

	df	Pseudo-F	P (perm)	Variance explained	df	Pseudo-F	P (perm)	Variance explained
	0–15 cm				15–30 cm			
Eukaryotes at class level - sM								
TIL[a]	1	2.24	**0.025**[b]	**11.96**	1	4.02	**0.003**	**22.69**
N fert	1	2.11	**0.017**	**10.76**	1	2.21	**0.012**	**9.05**
TIL x N fert	1	1.99	**0.024**	**19.13**	1	2.55	**0.019**	**23.20**
PERMDISP								
TIL			0.693				0.071	
N fert			0.078				0.359	
Eukaryotes at class level - mM								
TIL	1	4.60	**0.003**	**19.94**	1	1.05	0.411	–
N fert	1	3.90	**0.002**	**16.02**	1	1.64	0.127	–
TIL x N fert	1	3.79	**0.006**	**30.85**	1	0.90	0.569	–
PERMDISP								
TIL			0.237				–	
N fert			0.362				–	
Eukaryotes at phylum level - sM vs mM								
Matrix[c]	1	1.07	**.0.001**	**38.31**	1	11.18	**0.001**	**42.51**
TIL x N fert	1	2.90	**0.004**	**3.74**	1	1.22	0.287	4.51
N fert	1	1.54	0.140	10.69	1	2.36	**0.014**	**2.84**
PERMDISP								
Matrix			**0.025**				0.069	

[a] PERMANOVA was performed following a split-plot design with tillage (TIL) as main-plot factor and N fertilization (N fert) as subplot factor and with three replicate plots per treatment: TIL (minimum tillage and conventional tillage) and N fert (0 kg N ha^{-1} and 200 kg N ha^{-1}).
[b] In bold statistically significant values ($P \leq 0.05$).
[c] PERMANOVA was performed using the matrix as fixed factor, TIL and N fert as covaribales and 12 replicate plots per matrix.

detection of *Imbricatea* within sM under MTN0 supports that N availability is a major driver of cercozoan communities. This class had unexpectedly high abundance in sM under MTN0 (Fig. 1b), although it was shown to be highly favoured by organic fertilizers (Lentendu et al., 2014). According to our results (Fig. 1b), the heterotroph lineage Tubulinea are dominant in highly N-fertilizer soils (Sun et al., 2021), and the ciliate *Nassophorea*, characterizing the community structures of sM under MTN0, supports their role in energy transfer between trophic levels under low N availability (Gao et al., 2016). Finally, the dominance in sM under MT of *Enoplea* and *Chromadorea* known to be plant and animal nematode parasites is consistent with their general trend in soil aggregates (Jiang et al., 2017). Their abundance in N0 and N200, respectively, might be explained by specific predator-prey interactions occurring within intra-aggregate pores at differential N availabilities.

At the subsurface layer (Fig. 1c), N fertilization drove a stronger shift of the eukaryotic community structure in sM under CT compared with MT that showed similar structures irrespective to N fertilization, highlighting an opposite pattern as compared with the one observed at surface layer (Fig. 1b). Under CTN0, sM was characterized by high abundance of *Chilopoda* (Arthropoda, Animalia) and *Sarcomonadea*

(Cercozoa, Rhizaria), and under CTN200 by high abundance of *Colpodea* and *Conoidasida* (Alveolata) (Fig. 1c). Indeed, under ploughing, the subsurface layer is less compacted than under MT and shows a lower bulk density, resulting in an increase of pore size and aeration (Berisso et al., 2012). This may allow a larger root development under N fertilization and a higher variation in soil moisture and temperature compared with no fertilization (Piazza et al., 2020). Small macroaggregates are inaccessible to living centipedes (*Chilopoda*), thus their abundance in this fraction under CTN0 can be only related to a role as binding agents or as dead biomass consumed by decomposers. By contrast, *Sarcomonadea*, previously found in soil as the dominant class within the phylum Cercozoa (Degrune et al., 2019b), are likely to play an active role also in sM at low N availabilities. The abundance of *Colpodea* in CTN200 at the subsurface layer (Fig. 1c) is consistent with their dominance at the surface layer (Fig. 1b) and can be explained by bacterial pray changes following N application, while no information is available on *Conoidasida* trophic functional role.

In MTN0 and MTN200 at subsurface layer sM were characterized by *Chromadorea* (Nematoda), *Vampyrellidea* (Cercozoa), OTU1*Nassophorea* (Ciliophora, Alveolata) and many fungi (e.g., *Glomeromycetes*, *Sordariomycetes*) (Fig. 1c). The abundance of *Choromadorea* and *Nassophorea* is consistent with surface layer observations. Moreover, *Vampyrellidea,* observed for the first time in sM, are fungivorus that may control the parasitic rust fungus of wheat under MT (Adl and Gupta, 2006). Finally, the abundance of *Glomeromycetes* and *Sordariomycetes* confirms their crucial role in driving soil aggregates under undisturbed conditions (Rillig et al., 2015; Wang et al., 2021).

In mM, a strong interaction effect between tillage and N fertilization was found on the eukaryotic community structures only at surface layer, consisting in a strong shift of the structure under CTN0 compared with the other treatments (Table 1, Fig. 1d). This effect is in line with the aggregate distribution of mM found by Piazza et al. (2020). The shifts of aggregate distribution and eukaryotic community structure toward more mM and distinct soil biota communities under CT at low N availability can be related to a lower sequestration of C within mM and thus in differences of the related functional soil biota. By contrast, the lack of effect at subsurface layer is unexpected since the percentage of mM was significantly decreased by tillage intensification (CT < MT; - 21%) (Piazza et al., 2020). However, this inconsistency could be due to the coverage of the V4 region primer set, its taxonomic resolution or limitation in amplifying rare taxa or taxa with lower proportions of template DNA in DNA extracts (Choi and Park, 2020). Moreover, considering aggregate pore size and animal body size, the presence of traces of animal DNA (i.e., nematodes, Arthropoda) within aggregates is likely not attributable to the occurrence of living animals, but to the process of aggregate formation which utilises organic decaying material as binding agent.

Nitrogen fertilization determined a strong shift at the surface layer in the eukaryotic community structure of mM under CT (Fig. 1d) from *Glomeromycetes* (Glomeromycota), *Imbricatea*, *Sarcomonadea* and *Vampyrellidea* (Cercozoa) in CTN0 to OTU1*Xantho* (Ochrophyta), *Oligohymenophorea* (Ciliophora), Stramenopiles (Chromista) and the fungus *Tremellomycetes* in CTN200. This is the first time that Glomeromycota have been detected within mM fraction. Previously, using a cloning approach targeting the long-fragment SSU-ITS-LSU (Krüger et al., 2009) we could not detect AMF within mM (data not shown), and this was also supported by several works reporting their major roles only in macroaggregates (e.g., Miller and Jastrow, 2000; Rillig et al., 2002). However, in this study, the observed large proportion of Glomeromycota (14%) in mM under CTN0 supports the fact that tillage under unfertilized conditions may not negatively affect the development of the extraradical mycelium, potentially improving the production of glomalin and enhancing soil aggregate stability (Bedini et al., 2009). Similarly, the high abundance of Cercozoa in mM under CTN0 suggests for the first time that this phylum plays a major role within mM under ploughed and no fertilized conditions at surface layer. This is consistent with the findings of Degrune et al. (2019a) that highlighted under ploughing and at topsoil distinct cercozoan communities in microhabitats (i.e., drilosphere and rhizosphere) compared with bulk soil. Moreover, the distinct eukaryotic community found at surface layer in mM under CTN200 additionally supports that, under ploughing, nutrient availabilities in microhabitats allow the dominance of functional protists (Alveolata: *Oligohymenophorea*; Chromista: OTU1*Xantho* and Stramenopiles), potentially contributing to OM decomposition and mineralization through several functional groups. In addition, scarce information is available on the functional roles in agricultural soils of *Tremellomycetes*, a heterogeneous group comprising saprotrophs, animal parasites, and fungicolous species.

Similar community structures were observed at surface layer within mM in MTN0 and MTN200 [e.g., *Sarcomonadea*1 (Cercozoa), and fungi such as *Dothideomycetes* (Ascomycota) and OTU1Cystobasidio (Basidiomycota)] (Fig. 1d). These results support the hypothesis of a major role played by Cercozoa together with distinct classes of fungi also within mM under MT. However, it is well known that the 18S barcoding utilised in this work is less efficient compared with the ITS for detecting many groups of fungi (Schoch et al., 2012).

To test if soil aggregates differentially shape the structure of soil eukaryotes (hypothesis 2), the relative abundance pattern of taxa was determined in sM and mM. Significant differences among matrices (sM *vs* mM) were found and supported by PERMANOVAs (Table 1). PCO biplots showed that at both soil layers more phyla were linked to mM as compared with sM (Fig. 3c and d). Recently, Liao et al. (2018) and Wang et al. (2021) used an Illumina sequencing approach for studying at phylum and class level the bacterial and fungal community structures within soil aggregates across different fertilization treatments. Although differences in community structure were detected for both bacteria and fungi, fungal community in sM and free microaggregates differed more than bacteria (Liao et al., 2018; Wang et al., 2021). Our findings support that some unresolved taxa belonging to Glomeromycota and Ascomycota are positively associated with mM, as previously reported in free microaggregates for unclassified Ascomycota (Wang et al., 2021) and for a group of unclassified *Glomerales* (Lu et al., 2018). Similarly to the results of Jiang et al. (2017), the total abundance of nematodes increased with increasing aggregate size. Finally, the alveolate Apicomplexa, Ciliophora, and Dinoflagellata were preferentially found in mM, whereas the amoebozoan Conosa and Lobosa in sM. This result additionally confirms the functional role played by protists within microenvironments. Moreover, Mollusca in mM and Arthropoda and Anellida in sM at both soil layers can be considered as preferential binding agents for aggregate fractions.

3.5. Eukaryotic taxa predictors for soil structuring and C stocks

To test if some eukaryotic taxa are predictors for soil structuring and C stock (hypothesis 3) we utilised a multiple regression analysis that allowed to identify the eukaryotic taxa that were good predictors for the amount of sM and mM and their SOC content, irrespective of management and soil depth (Table S1). Specifically, *Microbotryomycetes* and Alveolata were moderately strongly related to the amount of sM, with *Microbotryomycetes* identified as best predictor. Similarly, Cercozoa and *Chytridiomycetes* were related to the amount of mM, with Cercozoa playing the major role. Moreover, *Microbotryomycetes*, Cercozoa and Alevolata were moderately related to SOC in sM, with *Microbotryomycetes* consistently found to be the best predictor. Finally, *Chytridiomycetes* and Cercozoa were moderately related to SOC in mM, with *Chytridiomycetes* the best predictor. Previously, Bach et al. (2018) identified bacterial and fungal indicators in free and large macroaggregates. However, it is the first time that *Microbotryomycetes* and *Chytridiomycetes* have been shown to be correlated with the amount of sM and mM and their SOC content, respectively. Both fungal classes correlated with the pattern of C-cycling enzymes and SOC content in bulk soil (Piazza et al., 2019), and their abundance was high in sM and

mM, respectively (Degrune et al., 2019b). *Microbotryomycetes* were also recently identified as good predictors of slow and passive SOC decomposition parameters (Hale et al., 2019). Our findings on the positive association of protists, Alveolata and Cercozoa, with the amount of sM and mM and their SOC content, support the multiple agroecological roles of protists found in bulk soil (Cavalier-Smith and Chao, 2003; Delgado-Baquerizo et al., 2020). Moreover, our results confirm previous works reporting that protists are shaped by pore size reduction and soil aeration (Berisso et al., 2012; Degrune et al., 2019a), features related to soil aggregates. Thus, we can assume that Alveolata are playing major role in soil, promoting sM formation and slowing down the decay of SOM within sM, while Cercozoa are crucial microorganisms in mM taking part to long-term sequestration and storing of SOC.

3.6. How eukaryotes are interlinked among each other and to prokaryotes in soil aggregates, and network traits predictor for soil structuring and C stocks

To test if eukaryotes form structured assemblages and distinctive networks in soil aggregates (hypothesis 4), and how eukaryotes are linked to prokaryotes (hypothesis 5), within- and cross-domain networks were built for sM and mM. At the surface layer within- and cross-domain networks were more complex in mM than in sM, whereas at the subsurface layer they did not vary (Table 2, Fig. 4, Fig. S6). In the cross-domain networks, both sM and mM showed a general trend toward a higher percentage of eukaryotes per subnetwork compared to prokaryotes at the surface layer compared with the subsurface layer (Table 2, Fig. 4b,d). Moreover, at both soil depth, in the sM cross-domain networks most of subnetworks were composed of both eukaryotes and prokaryotes, whereas in the mM cross-domain networks the subnetworks were half composed of eukaryotes and half of prokaryotes, and few subnetworks were mixed (Table 2). This is the first study that demonstrated that eukaryotes, components of soil biota communities usually studied separately, formed structured associations within each other and with prokaryotes in soil aggregates. Moreover, mM consistently had tighter connectivity compared with sM in both within- and cross-domain networks (Fig. 4, Fig. S6). This might be related to microhabitat conditions (i.e., wetter and more nutrient rich microhabitats) in mM that shift biotic interactions from facilitation to competition, leading to higher correlations between eukaryotic taxa or eukaryotic and prokaryotic taxa, as previously reported for fungi and bacteria in bulk soil and roots across land uses and agricultural managements with a gradient of nutrient availabilities (e.g., SOC, P levels) and pH (de Menezes et al., 2015; Banerjee et al., 2016, 2018; Wang et al., 2021). Other explanations could be a higher proportion of viable cells and spores and a lower niche heterogeneity (i.e., nutrients) in mM respect to sM, leading to tighter within- and cross-domain networks. Finally, a higher plant residue diversity in mM could also explain the within- and cross-domain mM network traits, as previously shown for plant community composition or host selectivity against microbial network complexity (Xiong et al., 2021).

In the within-domain networks of sM and mM and at both soil layers, Cercozoa were highly co-occurring in the subnetworks respect to the other eukaryotic phyla, as shown by the network traits (Table 2, Supplementary Results and Discussions 2). It is noteworthy the high percentage of taxa belonging to Cercozoa (21%) involved in the largest subnetworks, composed of 33, occurring at surface layer in the within-domain mM networks. In addition, while fungi, mainly Ascomycota and Basidiomycota, were highly co-occurring in the within-domain sM subnetworks at both soil layer, protists, as Ochrophyta and

Table 2

Traits of within- (eukaryotes) and cross-domain (eukaryotes - prokaryotes) network in macroaggregates (sM) and occluded microaggregates (mM) at 0–15 and 15–30 cm soil depth (for the network diagrams see Figs. 4 and S10).

Traits	sM 0-15	sM 15-30	mM 0-15	mM 15-30
Eukaryotes				
Number of nodes excluding singletons	37	33	46	39
Number of edges	29	25	44	31
Number of singletons	12	16	8	15
Number of dyads	3	6	1	6
Number of subnetworks	6	5	4	5
Mean nodes per subnetwork	5.17 ± 0.60	4.20 ± 0.49	11.0 ± 7.34	5.40 ± 1.03
Linkage density (complexity)	1.57 ± 0.11 a	1.51 ± 0.13	1.89 ± 0.14 b	1.56 ± 0.13
% Positive interactions	78.4	75.8	95.7	79.5
Modularity	3	5	4	5
Identity of phyla with a frequency within the subnetworks ≥10%	Basidiomycota 15.9% Cercozoa 16.8%	Ascomycota 20% Basidiomycota 10% Cercozoa 11.7%	Cercozoa 11.6% Ochrophyta 16.7%	Chlorophyta 12% Cercozoa 21.7% Stramenopiles 16.7%
Eukaryotes - Prokaryotes				
Total number of nodes excluding singletons	79	83	96	67
Number of eukaryotic nodes excluding singletons	37	38	49	30
Number of prokaryotic nodes excluding singletons	42	45	47	37
Number of edges	74	72	109	54
Number of singletons	22	18	6	36
Number of dyads	3	8	2	5
Number of subnetworks	7	4	8	11
Mean nodes per subnetwork	10.43 ± 4.43	8.38 ± 2.74	23.50 ± 19.50	5.18 ± 0.70
Linkage density (complexity)	1.87 ± 0.11 a	1.76 ± 0.11	2.26 ± 0.12 b	1.61 ± 0.11
Modularity	9	9	11	9
Percentage of eukaryotes per subnetwork	68.39 ± 12.26	50.79 ± 12.03	74.70 ± 14.61	39.01 ± 14.69
Percentage of prokaryotes per subnetwork	31.61 ± 12.26	49.21 ± 12.03	25.30 ± 14.61	60.99 ± 14.69
Number of subnetworks with only eukaryotes	2	1	2	4
Number of subnetworks with only prokaryotes	0	1	0	5
Number of mixed subnetworks	5	6	2	2

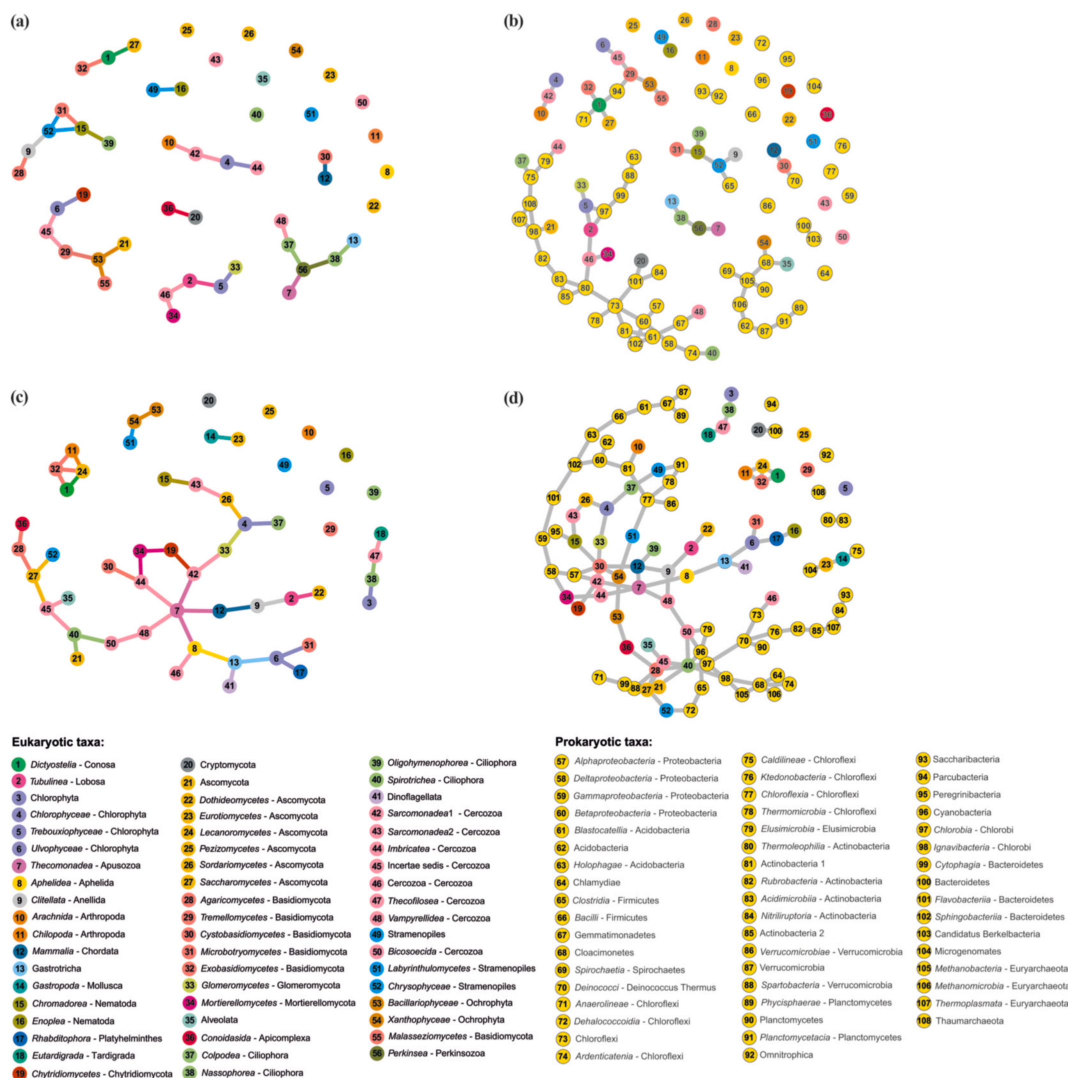

Fig. 4. Eukaryotic networks (within-domain networks) and eukaryotic-prokaryotic networks (cross-domain Associations networks) (at class phylogenetic resolution) in small macroaggregates (sM) (a,b) and occluded microaggregates (mM) (c,d) at surface layer (0–15 cm soil depth). Within-domain networks were built using the SParse InversE Covariance estimation for Ecological ASsociation Inference (SPIEC-EASI) package version 0.1 in R (https://github.com/zdk123/SpiecEasi/), while cross-domain networks were built by the cross-domain extension of SPIEC-EASI (Kurtz et al., 2015; Tipton et al., 2018). Details about network construction are given in Material and Methods and Supplementary Methods 3, and R scripts are provided in Supplementary Methods 4.

Stramenopiles together with Chlorophyta, were co-occurring in the within-domain mM subnetworks at surface and subsurface layer, respectively (Table 2). Following the definition of keystone taxa by Banerjee et al. (2018), we can support the agroecological theory that some Cercozoa, as *Sarcomondea* and *Vampyrillidea* (Fig. 4c), can be suggested as "hubs" (keystone taxa) within mM. Consistently, in the sM and mM cross-domain networks at the surface layer, Cercozoa can be considered as "hubs" as they are directly connected to prokaryotes in the largest subnetworks (Fig. 4b,d). Moreover, protists were involved in the largest subnetworks and showed variable direct connections (Fig. 4b,d, Figs. S6b and d). Indeed, the linkages of protists, according to their feeding versatility (Geisen, 2016), varied from direct connections to many prokaryotes, mainly belonging to Chloroflexi in the cross-domain sM networks (Fig. 4b, Fig. S6b), to direct linkages to other eukaryotes (e.g., fungi) or other protists in mM networks (Fig. 4d, Fig. S6d). So far, studies on the functional roles of soil biota in the formation of soil aggregates have mainly focused on the role played by a single functional group, e.g. AMF, earthworms, nematodes, termites and microarthropods (mites and collembolans) (e.g., Lee and Foster, 1991; Pulleman et al., 2005; Rillig and Mummey, 2006; Siddiky et al., 2012a,b; Zhang et al., 2016). However, only recently, Cercozoa and other protists, as Lobosa and Ciliophora, were shown to be positively related with ecosystem services, i.e. nutrient cycling and OM decomposition (Delgado-Baquerizo et al., 2020). Details about traits and taxa co-occurrences in within- and cross-domain networks and the significance of cross-domain relationships (Mantel test and CoCA) are given in Supplementary Results and Discussion 2, Fig. S7 and Fig. S8.

To test if some network traits can predict soil structuring and C stocks, we utilised a multiple regressions analysis that identified the number of edges and mean nodes per network as predictors for the amount of sM and mM and their SOC content, irrespective of management and soil depth (Table S2). Although network analysis is now largely applied to study soil biota co-occurrence and plant-microbe associations across different treatments (e.g., de Menezes et al., 2015; Farrer et al., 2019; Feng et al., 2019), little is known about soil biota networks within aggregates (Jiang et al., 2015, 2017) and few studies have dissected the predictable power of the network traits (topological properties) on ecosystem services. Jiang et al. (2015, 2017) indicated that aggregate fractions (large and small macroaggregates, and free microaggregates) showed a strong effect on the association networks of

nematodes and bacteria and using the topological properties they could identify large macroaggregates network as organized soil food web, showing functional interrelationships between bacterivorous nematodes and bacteria. In accordance with our results, the topological properties of soil biota networks should be taken into consideration for dissecting soil structuring as well as C cycling.

4. Conclusions

We have shown that soil aggregation is essential for a complete 'multifunctional' perspective of soil biota. A full understanding of relationships between soil biota and soil functions requires analyses emphasizing the feedbacks between soil structure and soil biota, rather than a unidirectional approach simply addressing the roles of single key functional groups. Next generation sequencing tools have been confirmed in this study to be crucial in the understanding of eukaryotic structures and soil biota networks and have the potential to further reveal their contributions to soil functions. Indeed, our findings demonstrate for the first time that protists together with fungi play major roles in soil structuring and C cycling, and that Cercozoa represent hubs in the soil biota aggregate networks. This supports the fact that their conservation is fundamental to prevent soil degradation and to enhance SOC accumulation in agroecosystems.

Declaration of competing interest

The authors declare that they have no known competing financial interests or personal relationships that could have appeared to influence the work reported in this paper.

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Appendix A. Supplementary data

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